SOIL SEED BANKS IN PLANTATIONS AND TROPICAL SEASONAL RAINFORESTS OF XISHUANGBANNA, SOUTH-WEST CHINA

H Chen^{1, 2}, M Cao^{1, *} & Y Tang¹

¹Key Laboratory of Tropical Forest Ecology, Xishuangbanna Tropical Botanical Garden, Chinese Academy of Sciences, Kunming 650223, PR China
²University of Chinese Academy of Sciences, Beijing 100049, PR China

Received May 2012

CHEN H, CAO M & TANG Y. 2013. Soil seed banks in plantations and tropical seasonal rainforests of Xishuangbanna, South-West China. The impacts of land conversion from forests to plantations on seed availability are poorly appraised in terms of future restoration potentials. Seed storage and species composition of soil seed banks were characterised in five natural forests and three plantations. Mean soil seed storage differed significantly between study sites, ranging from 2444 seeds/m² in a selectively logged forest to 17,946 seeds/m² in a pomelo orchard. For all sites except for secondary forest 2, herb seeds dominated the soil seed bank. More seeds of pioneer tree species were found in the soil seed banks of secondary forests than in primary rainforests. This suggests that secondary forests have higher seed availability and stronger potential for forest restoration than primary forests. Anthropogenic management activities resulted in abundant but less diverse soil seed bank due to the dominance of a few herbaceous species in the pomelo orchards and rubber plantations. Low tree seed storage was found in soil seed banks of pomelo orchards and rubber plantations and might decelerate forest succession. Thus, the seeds of pioneer trees in soil seed banks of secondary forests could promote natural forest regeneration in abandoned plantations.

Keywords: Invasive herbs, management, pioneer tree species, seed availability, tropical forest

CHEN H, CAO M & TANG Y. 2013. Kebolehdapatan biji benih di ladang dan hutan hujan tropika bermusim di Xishuangbanna, Barat Daya China. Kesan penerokaan hutan untuk dijadikan ladang kurang dikaji dari segi kebolehdapatan biji benih bagi pemulihan hutan. Simpanan biji benih dan komposisi spesies di tanah asal lima hutan semula jadi dan tiga ladang dikaji. Purata simpanan biji benih berbeza dengan ketara antara tapak kajian dan berjulat antara 2444 biji benih/m² di hutan yang mengalami tebangan memilih hingga 17,946 biji benih/m² di kebun limau bali. Semua tapak kajian kecuali hutan sekunder 2 didominasi oleh biji benih tumbuhan herba. Lebih banyak biji benih spesies pokok perintis ditemui di tapak hutan sekunder berbanding dengan hutan primer. Ini mencadangkan yang hutan sekunder mempunyai sumber biji benih yang lebih banyak. Oleh itu, potensi hutan sekunder untuk memulihkan hutan adalah lebih baik berbanding dengan hutan primer. Aktiviti urusan antropogeni membanyakkan biji benih tetapi mengurangkan kepelbagaian biji benih disebabkan dominasi beberapa spesies tumbuhan herba di kebun limau bali dan ladang getah. Simpanan biji benih yang rendah di kebun limau bali dan ladang getah mungkin melambatkan sesaran hutan. Oleh itu, biji benih pokok perintis di tapak hutan sekunder boleh digunakan sebagai sumber untuk memulihkan hutan semula jadi di tapak ladang yang terbiar.

INTRODUCTION

Tropical ecosystems are exceptionally rich and exclusive reservoirs of much of the biodiversity on earth (Whitmore 1990). However, human population and industrial timber logging have led to rapid and extensive destruction of tropical forests (Laurance 2007). The conversion of rainforests into secondary forests or agricultural plantations is occurring at an alarming rate. Over the next 50 years, rapid and widespread agricultural expansion poses a serious threat to natural ecosystems worldwide (Tilman et al. 2001). The highest rate of deforestation in any major tropical region occurs in South-East Asia, which can lose three quarters of its original

^{*}caom@xtbg.ac.cn

forests and up to 42% of its biodiversity by 2100 (Sodhi et al. 2004). Furthermore, the planting of perennial exotic crops, which include rubber (*Hevea brasiliensis*), oil palm (*Elaeis guineensis*) and coconut (*Cocos nucifera*), accounts for 20–30% of the total cultivated area of South-East Asia and has led to extensive deforestation (Flint 1994).

Xishuangbanna $(1.96 \times 10^6 \text{ ha})$ is famous in China for being a biodiversity hotspot with diverse tropical flora and fauna (Zhang & Cao 1995, Zhu et al. 2006). However, the forest cover of Xishuangbanna declined from 69% in 1976 to less than 50% in 2003 because of rapid development of the local economy leading to increased demand for natural resources (Li et al. 2007, 2009). Since reforms and clearing of lands in the late 1970s, traditional agriculture such as slash-and-burn cultivation has been replaced by permanent cash crop cultivation. The development of rubber plantations is a major threat to local tropical rainforests, which are usually distributed below 900 m above sea level in Xishuangbanna. From 1976 till 2003, a total of 139,576 ha of tropical rainforests were converted into rubber plantations at lower altitudes. Furthermore, during the last decade, the rising market price of rubber further increased its demand and caused more severe clearance of natural forests distributed at higher altitudes or steep slopes for new arable land (Li et al. 2007, 2009). Meanwhile, pomelo (Citrus maxima) orchards expanded rapidly to meet market demand. The change in landuse from natural forests to plantations causes not only biodiversity loss but also deterioration of soil seed banks (Lemenih & Teketay 2006).

Soil seed bank is the aggregation of viable seeds in the superficial soil (Simpson et al. 1989). Seed inputs into the soil seed bank are determined by seed rain from the standing vegetation (Guevara 2005). Both the species diversity and size of soil seed bank are reduced in primary forests compared with secondary forests because of the depletion and limited seed inputs of pioneer trees (Plue et al. 2010). Natural or anthropogenic disturbances of standing vegetation might also affect the seed density and species composition of soil seed bank (Mamede & de Araujo 2008). Extensive and frequent disturbances lead to larger soil seed bank, which is due to large inputs of seeds from light-demanding and small-seeded plant species (Alvarez-Aquino et al. 2005). Viable pioneer tree seeds in soil seed bank can promote forest regeneration after natural or anthropogenic disturbances (Zimmerman et al. 2000). In contrast, a lack of pioneer tree seeds in soil seed bank impedes the secondary succession of indigenous forest in abandoned pastures (Wijdeven & Kuzee 2000, Zimmerman et al. 2000). Therefore, seed availability, which hereafter refers to the viable seed storage of pioneer tree species in soil seed bank, serves as a key indicator for the assessment of the restoration potential of vegetation. Ex-situ transplant of soil seed bank that contains seeds of many pioneer trees is often a recommended low-cost technique for restoration of vegetation that is degraded by mine tailings (Parrotta & Knowles 2001, Zhang et al. 2001, Mackenzie & Naeth 2010).

In Xishuangbanna, studies have revealed higher seed storage in a seasonal rainforest and a dipterocarp rainforest than in some other rainforests of tropical Asia, reflecting the effects of local forest fragmentation (Cao et al. 2000, Tang et al. 2006). In addition, a majority of nonnative species, especially wind-dispersed weeds, comprises the soil seed bank of forest-edge areas and is induced by human disturbance (Lin et al. 2006). Although slash-and-burn agriculture eliminates most of the superficial seeds in the soil seed bank, abundant seed storage (65,665 seeds/m²) was found in young slash-andburn fallow (Cao et al. 2000).

Based on previous findings, the present study was conducted to examine the restoration potential of different types of lands by comparing the seed availability of soil seed bank. We hypothesise that secondary forests have more abundant storage of pioneer tree seeds than primary rainforests or plantations or pomelo orchards, thereby resulting in relatively high seed availability in its soil seed bank. Reduced seed availability in soil seed bank is a main barrier to forest recovery in rubber plantations and pomelo orchards. The reduced availability is due to depletion of tree seeds stored in the soil and predominance of exotic/non-native herbs after long-term plantation practices. The objectives of this study were to (1) investigate the changes in species composition and seed density of soil seed bank following deforestation and conversion into plantations and (2) examine the natural restoration potentials of sites by comparing the seed availability in the soil seed bank. We also presented some perspectives on the promotion of forest restoration.

MATERIALS AND METHODS

Study area

This study was conducted in Xishuangbanna, South-West China, where tropical seasonal forests are distributed (Cao et al. 2006). The climate is tropical monsoon with annual mean temperature of 21.7 °C. Average annual rainfall is approximately 1500 mm and more than 80% of it occurs during the rainy season from May till October. During the dry season, especially from November till February, there is high frequency of heavy fog in the lower hills and valleys in the night and morning (Liu et al. 2004, 2005, Cao et al. 2006). The primary vegetation is tropical seasonal rainforest that consists of Pometia tomentosa, Terminalia myriocarpa, Barringtonia macrostachya, Shorea wantianshuea, Gironniera subaequalis and Chisocheton siamensis (Zhu et al. 2006).

A total of eight forest types were chosen that represented different levels of disturbances. They were two seasonal rainforests (rainforest 1, RF1; rainforest 2, RF2), one selectively-logged rainforest (SLF), two secondary forests (secondary forest 1, SF1; secondary forest 2, SF2), one Senna siamea plantation (SSP), one monoculture rubber plantation (RUP) and one monoculture pomelo orchard (POO) (Table 1). The rainforest 1 and rainforest 2 sites are located in the core zones of Menglun and Mengla Nature Reserves respectively, where collection and hunting are forbidden. The other six sites are located in the Xishuangbanna Tropical Botanical Garden (XTBG). The selectively logged rainforest is a fragmented, tropical seasonal rainforest. It was selectively logged in 1959. Since then, it has been protected as an in-situ reserve for indigenous plant species. Secondary forest 1 and secondary forest 2 are secondary forests that naturally regenerated after slash-and-burn cultivation approximately 30 years ago. The S. siamea plantation is a fuelwood plantation that was established by clearing the secondary forest 30 years ago. Since XTBG assumed its management approximately 15 years ago, fuelwood harvest has been forbidden. The rubber plantation and pomelo orchard sites were planted after slashing and burning of secondary forests. Intensive management activities, which include manual weeding or three herbicide (glyphosate) applications during the rainy season to eliminate weeds, are conducted in the rubber plantation and pomelo orchard every year. In August, a compound fertiliser containing N, P and K was applied under each rubber tree in the rubber plantation. At the end of the rainy season, the farmers fertilise the pomelo orchard. In addition, the rubber plantation site has rubber latex tapped once every 2 days in April till November of each year. In the pomelo orchard site, fruits are harvested from July till August of each year.

Soil sampling

In May 2009, 20 soil samples (diameter 10 cm, depth 5 cm) were collected using a steel tube at 10 m intervals along a 200 m transect in the core area of each site. The surface area of the soil sampled at each stand was 78.5 cm² and the soil volume was 392.7 cm³. A total of 160 soil samples were collected from the eight sites.

Parameter	RF1	RF2	SLF	SF1	SF2	SSP	RUP	POO
Altitude (m)	700	850	566	630	640	620	600	590
Latitude N	21° 57'	21° $42'$	21° 55'	$21^{\circ} 54'$	$21^{\circ} 54'$	$21^{\circ} 54'$	$21^{\circ} 54'$	$21^\circ~54'$
Longitude E	$101^\circ 12'$	101° 35'	$101^{\circ} 35'$	$101^\circ \ 16'$	$101^\circ~16'$	$101^\circ \ 16'$	$101^\circ~16'$	$101^{\circ} 16'$
Area (ha)	CF	CF	100	14	8.7	15.7	10.8	22.3
Canopy height (m)	35	40	30	16	16	18	15	3
Age (years)	> 200	> 200	> 200	30	30	30	19	15
Disturbance	Low	Low	Low	Medium	Medium	Medium	High	High

 Table 1
 Characteristics of study sites

RF1 = rainforest 1, RF2 = rainforest 2, SLF = selectively logged rainforest, SF1 = secondary forest 1, SF2 = secondary forest 2, SSP = *Senna siamea* plantation, RUP = rubber plantation, POO = pomelo orchard, CF = continuous forest

Seed germination

The number of viable seeds in each soil sample was estimated through germination trials. We refer to viable seeds as those germinated under our germination condition settings. Each soil sample was placed in a separate tray at a depth of approximately 1 cm in a non-temperature controlled greenhouse (30% of full light). The trays were watered daily to keep the soil samples moist. Seedlings were counted and identified seedlings were removed. Seedlings that were difficult to distinguish were transplanted into pots for continued growth until identification was possible. After emerged seedlings were removed from the trays, soil samples were carefully stirred. Trays were kept and observed until no seedlings had emerged for a period of 1 month. There might be some seeds dormant in the soil, even at the end of the germination trial. However, we might also germinate seeds that persisted in the soil for a long time and those which dormancy was broken during our germination trials. These serve as mutual complementation in theory.

Data analysis

Frequency at a site was defined as the percentage of soil samples with an identified species in the soil seed bank of that site. Seed abundances between sites were compared by one-way analysis of variance using log (n + 1) transformations where n = number of seedlings germinated (Tang et al. 2006). The proportions of tree seeds to herb seeds were tested for differences between sites using Pearson chi-squared tests.

RESULTS

Seed stocks and plants

A total of 6517 seeds germinated from the soil seed banks of the eight sites (Table 2). Mean seed density (number of germinated seeds/m²) in the top 5 cm of soil varied significantly from 2445 \pm 975 seeds /m² (mean \pm SD) in the selectively logged rainforest to 17,946 \pm 8366 seeds /m² in the pomelo orchard (Figure 1) (p < 0.05). Seed density of the soil seed bank was significantly higher in the pomelo orchard and secondary forest 2 than the other sites. Only 39 and 41





species germinated from soil samples of rubber plantation and pomelo orchard respectively, whereas at least 50 species germinated from each of the other six sites (Table 2).

Herb seeds dominated the soil seed banks of all the sites except for secondary forest 2, which was dominated by tree species (Table 2). Herbs represented more than 82.1% of the species and 95.9% of the plants found in the rubber plantation and pomelo orchard. However, the proportions of woody (tree + shrub) to herbaceous seeds in the soil seed bank were significantly smaller in the rubber plantation and pomelo orchard than in other sites. No vine seeds germinated from the soil samples of the rubber plantation or pomelo orchard. In the S. siamea plantation and in five forest sites except for secondary forest 2, woody species accounted for less than half of the seeds. Secondary forest 2 had the most tree seeds followed by secondary forest 1, S. siamea plantation and rainforest 2. Although there were fewer tree seeds stored in the soil seed banks of rainforest 1 and selectively logged rainforest, species diversity represented in the soil seed bank was comparable with those found in the soil seed banks of secondary forest 1 and secondary forest 2 (Table 2).

Site		Seed	composition				Spe	ecies compositio	n	
I	Tree	Shrub	Herb	Vine	Total	Tree	Shrub	Herb	Vine	Total
RFI	76 (19.3%)	81 (20.5%)	230 (58.2%)	8 (2.0%)	395	15(28.3%)	4 (7.6%)	30 (56.6%)	4 (7.5%)	53
RF2	90 (17.5%)	35(6.8%)	370~(71.8%)	20 (3.9%)	515	$14\ (28.0\%)$	9(18.0%)	24 (48.0%)	3 (6.0%)	50
SLF	80 (21.0%)	57~(14.9%)	232 (60.7%)	13 (3.4%)	382	$14\ (28.0\%)$	6(12.0%)	25 (50.0%)	5(10.0%)	50
SF1	127~(23.7%)	81 (15.2%)	318~(59.4%)	9(1.7%)	535	13 (24.5%)	10 (18.9%)	26(49.1%)	4 (7.5%)	53
SF2	341 (38.2%)	209 (23.4%)	323 (36.2%)	20 (2.2%)	893	15(27.3%)	8 (14.5%)	26 (47.3%)	6(10.9%)	55
SSP	101 (25.5%)	52~(13.1%)	219~(55.3%)	24(6.1%)	396	$8\ (13.1\%)$	9(14.8%)	33 (54.1%)	11 (18.0%)	61
RUP	9(1.5%)	15(2.6%)	558~(95.9%)	(%0) 0	582	3 (7.7%)	4 (10.3%)	32 (82.1%)	(0%) (0%)	39
000	10 (0.4%)	3 (0.1%)	2806 (99.5%)	(0%) = 0	2819	3 (7.3%)	2(4.9%)	36 (87.8%)	(%0) (0%)	41

 Table 2
 Plant and seed compositions in the soil seed banks of the eight study sites

RF1 = rainforest 1, RF2 = rainforest 2, SLF = selectively logged rainforest, SF1 = secondary forest 1, SF2 = secondary forest 2, SSP = *Senna siamea* plantation, RUP = rubber plantation, POO = pomelo orchard

Pioneer tree species

There were more pioneer tree seeds in secondary forest 1 and secondary forest 2 than in rainforest 1, selectively logged rainforest and rainforest 2 (Figure 2a). Only three tree species were observed in the soil seed bank of rubber plantation (Figure 2b). They were Ficus hirta, F. hispida and Macaranga denticulata (Table 3). Similarly, we found only three tree species in the soil seed bank of pomelo orchard and they were Alchornea davidii, Trema orientalis and F. hirta. All tree seeds in the soil seed banks of rubber plantation and pomelo orchard were pioneer tree species. Trema orientalis dominated the tree species recorded in the soil seed banks of the eight sites with a total of 281 seeds. A total of 215 T. orientalis seeds were found in the soil seed banks of secondary forest 1 and secondary forest 2. Ficus was the most abundant genus in the eight sites with a total of 396 seeds belonging to nine species. In all sites, F. hirta was the only tree species found in the soil seed banks. Bird-dispersed seeds of T. orientalis, *M. denticulata* and *Ficus* spp. accounted for more than 80% of the animal-dispersed seeds of woody species in all sites. The only two wind-dispersed woody species, Duabanga grandiflora (15 seeds) and Tetrameles nudiflora (11 seeds), were found in the soil seed banks of the rainforest 1 and selectively logged rainforest.

Exotic herbs

The pomelo orchard had the most exotic herb seeds, followed by rubber plantation, secondary forest 2, secondary forest 1, S. siamea plantation, rainforest 1, selectively logged rainforest and rainforest 2 (Figure 3a). The most abundant exotic herbs were Ageratum conyzoides, Crassocephalum crepidioides, Spermacoce latifolia, Convza sumatrensis and Chromolaena odorata with 1041, 223, 190, 155 and 143 seeds respectively (Table 4). The less diverse exotic herbs accounted for nearly half of the herb seeds stored in the secondary forest 2, rubber plantation and pomelo orchard, whereas the diverse native herb seeds accounted for more than 70% of the herb seeds stored in the other sites (Figures 3a and b). Although herbs dominated the soil seed banks of rubber plantation and pomelo orchard, only C. crepidioides, C. sumatrensis, C. odorata, Vernonia cinerea and Stevia rebaudiana were wind-dispersed, and all of these species belong to the Asteraceae. Furthermore, 116 (result not



Figure 2 Tree seed storage as determined by (a) seeds and (b) species in the eight sites (mean ± SD); different letters indicate significantly different sites at p < 0.05; RF1 = rainforest 1, RF2 = rainforest 2, SLF = selectively logged rainforest, SF1 = secondary forest 1, SF2 = secondary forest 2, SSP = *Senna siamea* plantation, RUP = rubber plantation, POO = pomelo orchard

shown) and 159 (Table 4) wind-dispersed seeds were found in the soil seed banks of the rubber plantation and pomelo orchard respectively.

DISCUSSION

Conversion of rainforest to secondary forests

The results of our study confirmed that seed density was significantly higher at earlier stages of succession than in later stages. More seeds were stored in the soil seed banks of secondary forest 1 and secondary forest 2 than rainforest 1 and selectively logged rainforest (Table 2 and Figure 1). Although relatively high numbers of seeds (2514 to 3284 seeds/m²; results not shown) were recorded in this study in seasonal rainforests

Sheries	Family	Functional	Seed	RFI	RF9	SIF	SF1	SF9	SSP	RUP	DOO	Total	Frequency
	(group	dispersal		1		5	1	1	1041			humbers
Trema orientalis	Ulmaceae	Pioneer	Animal	3	21	6	26	189	30	0	3	281	0.875
Ficus nervosa	Moraceae	Pioneer	Animal	19	4	8	35	62	19	0	0	147	0.75
Ficus hirta	Moraceae	Pioneer	Animal	4	14	3	14	20	7	1	5	65	1.000
Ficus langkokensis	Moraceae	Shade-tolerant	Animal	7	19	20	5	9	11	0	0	65	0.75
Macaranga denticulata	Euphorbiaceae	Pioneer	Animal	0	1	0	23	9	19	1	0	50	0.625
Ficus hispida	Moraceae	Pioneer	Animal	0	ы	4	7	17	13	7	0	46	0.75
Ficus auriculata	Moraceae	Shade-tolerant	Animal	24	1	Г	7	0	0	0	0	33	0.500
Euodia trichotoma	Rutaceae	Shade-tolerant	Animal	5	6	0	0	20	0	0	0	31	0.375
Ficus semicordata	Moraceae	Pioneer	Animal	ы	4	7	3	3	0	0	0	19	0.625
Ficus cyrtophylla	Moraceae	Pioneer	Animal	33	9	4	4	0	0	0	0	17	0.500
Duabanga grandiflora	Sonneratiacea	Pioneer	Wind	9	0	6	0	0	0	0	0	15	0.250
Mallotus paniculatus	Euphorbiaceae	Pioneer	Animal	1	5	1	3	9	1	0	0	14	0.750
Tetrameles nudiflora	Tetramelaceae	Shade-tolerant	Wind	1	0	10	0	0	0	0	0	11	0.250
Alchornea davidii	Euphorbiaceae	Pioneer	Others	0	0	1	1	3	0	0	5	6	0.500
Aralia armata	Araliaceae	Pioneer	Others	0	4	61	1	0	0	0	0	6	0.500
Rhus chinensis	Anacardiaceae	Pioneer	Animal	0	0	0	0	5	0	0	0	5	0.125
Ficus sp.	Moraceae	Pioneer	Animal	0	6	0	0	0	0	0	0	5	0.125
Ficus fistulosa	Moraceae	Shade-tolerant	Animal	0	0	0	0	0	0	0	0	6	0.125
Glochidion assamicum	Euphorbiaceae	Unknown	Others	1	0	0	1	0	0	0	0	5	0.250
Garuga floribunda var. gamblei	Burseraceae	Shade-tolerant	Others	1	0	0	0	0	1	0	0	5	0.250
Sauravia macrotricha	Saurauiaceae	Shade-tolerant	Animal	0	1	0	0	0	0	0	0	1	0.125
Maesa ramentacea	Myrsinaceae	Pioneer	Animal	1	0	0	0	0	0	0	0	1	0.125
Garuga pinnata	Burseraceae	Shade-tolerant	Others	0	0	0	0	1	0	0	0	1	0.125
Alangium barbatum	Alangiaceae	Shade-tolerant	Others	1	0	0	0	0	0	0	0	1	0.125
Canthium parvifolium	Apocynaceae	Shade-tolerant	Others	0	0	-	0	0	0	0	0	1	0.125
Others = gravity, ballistic and wate SSP = <i>Senna siamea</i> plantation, RU	er; RF1 = rainforest 1 JP = rubber plantatic	, RF2 = rainforest 2, n, POO = pomelo o	SLF = selecti rchard	vely log	ged rai	nforest	, SF1 =	second	lary fo:	rest 1, S	SF2 = se	condary	forest 2,



Figure 3 Herb seed storage as determined by (a) seeds and (b) species in the eight sites (mean ± SD); different letters indicate significantly different sites at p < 0.05; RF1 = rainforest 1, RF2 = rainforest 2, SLF = selectively logged rainforest, SF1 = secondary forest 1, SF2 = secondary forest 2, SSP = Senna siamea plantation, RUP = rubber plantation, POO = pomelo orchard

compared with other forest sites in South-East Asia (Metcalfe & Turner 1998, Singhakumara et al. 2000), our results were comparable with other studies on soil seed bank in Xishuangbanna (Cao et al. 2000, Tang et al. 2006). The abundant seed storage of rainforest 1 and rainforest 2 might be attributed to the highly fragmented distribution of local seasonal rainforests. However, the soil seed banks of rainforest sites had fewer seeds than the soil seed banks of secondary forests and plantations surveyed in this study. A study on soil seed bank at different stages of succession in Xishuangbanna also found higher seed density and greater number of pioneer tree seeds in the secondary forest than in the rainforest (Cao et al. 2000).

Higher soil seed storage densities found in earlier phases of forest succession are due to higher fecundity and early reproduction associated with greater seed longevity of pioneer species than primary species, which predominate in secondary forest sites (Dalling & Brown 2009). Sites that were surrounded by extensively disturbed areas such as plantations and pastures had high soil seed storage densities and contained large numbers of seeds of light-demanding shrubs and herbaceous species because of the seed inputs of pioneer species (Quintana-Ascencio et al. 1996). The secondary forest 1 and secondary forest 2 in present study were naturally regenerated after slash-and-burn cultivation. Many pioneer species then germinated from the soil seed bank and newly dispersed seeds from nearby forest because no more farming practice took place in the sites since they were abandoned.

Conversion of natural forest to plantation

Our study showed significant differences between the soil seed bank compositions of the rubber plantation and pomelo orchard with the soil seed bank compositions of other sites. Seeds of herbaceous species dominated the soil seed banks of the rubber plantation and pomelo orchard in terms of seed density and species number. These sites were intensively affected by human activities which eliminated tree seeds (Tables 2 and 3), although they significantly increased the seed abundance of herbs (Tables 2 and 4). Large soil seed banks of herbs were also observed in other studies. In Nigeria, herbaceous species made up 98% of the total seeds stored in citrus, oil palm, cashew and cocoa plantations; in contrast, only three woody species emerged from the soil seed bank of four plantations (Oke et al. 2007). In Mexico, only a few seeds of late successional tree species (< 2% of the species) were observed in the diverse soil samples that ranged from milpa fields to tropical rainforests (Quintana-Ascencio et al. 1996). Tree seedlings were continuously removed by repeated weeding in agricultural ecosystems. Thus, herbs and herbaceous seeds quickly dominated the seedling and soil seed bank respectively because of the short life cycle and high reproductive capacity of herbs. These changes eventually alter the soil seed banks composition, resulting in an abundance of herb seeds rather than seeds of successional woody species. These findings seem to suggest a deterministic role of intensive management in shaping soil seed banks in agricultural farming systems.

	sites
-	eight
5	ot the
-	anks (
-	Ω
-	d
	see
:	SOI
5	l the
	Ξ
-	cored
-	eeds si
-	erb s
-	ğ
ç	G
	species (
	loundant
	4 most 2
Ē	I ne z
	lable 4

Species	Family	Functional group	Seed dispersal	RF1	RF2	LSF	SF1	SF2	SSP	RUP	POO	Total	Frequency
Ageratum conyzoides	Asteraceae	Exotic	Others	2	4	0	4	ъ	6	169	845	1041	0.875
Torenia flava	Scrophulariaceae	Native	Others	18	0	19	19	0	9	4	720	788	0.875
Lindernia stricta	Scrophulariaceae	Native	Others	53	226	109	19	55	18	14	11	505	1
Saxifraga atolonifera	Saxifragaceae	Native	Others	0	0	0	37	44	5	5	226	314	0.625
Gonostegia hirta	Urticaceae	Native	Animal	0	0	0	0	0	1	5	290	296	0.375
Crassocephalum crepidioides	Asteraceae	Exotic	Wind	20	21	17	26	25	18	42	54	223	1
Thysanolaena maxima	Poaceae	Native	Others	39	42	4	50	28	38	4	5	213	1
Isachne globosa	Poaceae	Native	Others	4	0	0	49	6	9	132	4	204	0.75
Spermacoce latifolia	Rubiaceae	Exotic	Animal	0	0	0	29	8	4	9	143	190	0.625
Conyza sumatrensis	Asteraceae	Exotic	Wind	6	5	13	12	10	6	26	71	155	1
Chromolaena odorata	Asteraceae	Exotic	Wind	3	5	6	11	98	11	1	5	143	1
Hedyotis corymbosa	Rubiaceae	Native	Others	1	3	0	1	7	x	35	58	113	0.875
Vernonia cinerea	Asteraceae	Native	Wind	6	11	12	9	10	23	27	5	103	1
Kyllinga monocephala	Cyperaceae	Native	Others	0	1	1	9	0	1	9	61	78	0.875
Euphorbia hirta	Euphorbiaceae	Exotic	Others	25	3	5	0	0	0	61	31	99	0.625
Poaceae sp. 1	Poaceae	Native	Others	0	4	0	0	0	15	1	46	99	0.5
Cardamine flexuosa var. debilis	Brassicaceae	Native	Others	0	0	0	0	0	0	0	49	49	0.125
Phyllanthus urinaria	Euphorbiaceae	Native	Animal	1	0	0	0	0	0	4	35	42	0.5
Poaceae sp. 2	Poaceae	Native	Others	0	0	1	4	0	6	21	1	39	0.625
Cynoglossum amabile	Boraginaceae	Native	Others	16	0	0	6	0	11	0	0	38	0.5
Dichondra repens	Convolvulaceae	Native	Others	0	0	0	0	1	0	0	34	35	0.25
Stevia rebaudiana	Asteraceae	Exotic	Wind	0	0	0	0	0	1	8	24	35	0.5
Digitaria chinensis	Poaceae	Native	Others	0	0	3	0	${\mathfrak S}$	12	0	11	31	0.625
Ludwigia prostrata	Onagraceae	Native	Others	0	-	0	-	0	0	0	25	27	0.375
Others = gravity, ballistic and wa forest 2, SSP = <i>Senna siamea</i> plan	ater; RF1 = rainforest 1 ntation, RUP = rubber	, RF2 = rainfor plantation, PO	est 2, $SLF = 0$	selectiv orcha	'ely log rd	ged ra	infore	st, SF1	= seco	ondary	forest 1,	SF2 = se	condary

Pioneer tree species

Rapid growth and copious widely-dispersed seeds increase the probability of pioneer species reaching and settling in high light habitats. However, seedlings of shade-tolerant species experience rapid and high mortality because of reduced canopy shade (Swaine & Whitmore 1988). Forests with intermediate levels of disturbance have large soil seed banks dominated by pioneer tree species (Alvarez-Aquino et al. 2005). One of the most important tree species in the soil seed banks of secondary forest 1 and secondary forest 2 was T. orientalis. It is a shortlived pioneer tree species in the tropical Asia (Whitmore 1990). This tree species might even establish mono-dominant stands in abandoned, slash-and-burn fields (Cao & Zhang 1996). Trema seeds require disturbance of the soil surface and low canopy cover to increase light transmittance and stimulate germination respectively (Dalling et al. 1998). Therefore, Trema seeds may be stored in the soil seed bank under tropical forest canopies. Furthermore, dormant seeds lead to prolonged seed persistence of Trema in the soil (Dalling & Brown 2009). The most abundant tree group recorded in our study was fig with 9 species and 396 seeds. Fig seeds are photoblastic (Vázquez-Yanes et al. 1996). Therefore, these seeds may be stored in the soil seed bank because of poor light filtration through superficial soil (Bliss & Smith 1985).

Exotic herbs

Exotic herbs are thought to be super competitors because of their lack of enemies, high resourceuse efficiency and allelopathy (Callaway & Maron 2006, Funk & Vitousek 2007). Although herbs may not inhibit the settlement of tree seedlings (Aide & Cavelier 1994), several studies have shown that tree seedlings perform better after the clearance of herbs (Holl et al. 2000, Hooper et al. 2005). One of the most competitive exotic herbs observed in the present study was C. odorata. It is a perennial herb capable of forming dense canopy in degraded areas of the pan tropical region (Derouw 1991). Invasion of C. odorata was suggested as the primary factor responsible for poor regeneration of degraded forests in Ghana (Honu & Dang 2000).

Exotic herbs commonly occur in farmland and pastures. Exotic herbs need light for seed germination and seedling growth and can be excluded from the understorey by a closed forest canopy (Hooper et al. 2005). Most of the seeds stored in soil seed bank are dispersed on-site (Young et al. 1987). Thus, the seed inputs of exotic herbs were much higher in the rubber plantation and pomelo orchard than in the other forest sites, resulting in the soil seed banks of the rubber plantation and pomelo orchard containing more seeds of exotic herbs (Table 4).

Seed availability and restoration potential

High seed storage of pioneer tree species represents higher restoration potential. The enrichment of pioneer species in the soil seed banks may accelerate the rate of secondary succession (Zimmerman et al. 2000). Pioneer trees are adapted to degraded environments and their seed germination can be induced by high light or fluctuating temperature created by disturbance in forests (Pearson et al. 2002). In contrast, seeds of weeds dominating soil seed banks in intensively-managed agricultural fields showed low restoration potential (Wijdeven & Kuzee 2000, Zimmerman et al. 2000). Low seed availability is one of the main barriers to tropical rainforest restoration after intensive disturbance (Standish et al. 2007). Therefore, we recommend that quantitative assessment on restoration potential be conducted by examining seed availability of pioneer tree species and competitive herb seeds in soil seed banks.

The numbers of pioneer tree and exotic herb seeds in in the soil seed bank were higher in secondary forests than in primary forests. Furthermore, there were roughly equal numbers of pioneer tree and exotic herb seeds in primary forests. Nevertheless, the seed storage of pioneer trees was higher than that of exotic herbs in secondary forests, especially in secondary forest 2. Thus, the soil seed banks of secondary forests might promote early forest restoration. Many exotic herbs and few pioneer trees emerged from the soil seed banks of the rubber plantation and pomelo orchard, suggesting that natural regeneration might be arrested in these areas.

CONCLUSIONS

Soils from secondary forests had larger numbers and more species of pioneer tree seeds than rainforests. The conversion of natural forest into rubber plantations and pomelo orchards had exhausted stored seeds of tree species in the soil and increased the seed density of exotic herbs in the soil seed bank. Thus, there was low seed availability for forest restoration in the intensively-managed plantations and orchards. However, the addition of soil seed banks from secondary forests, which contained abundant seeds of pioneer tree species in abandoned plantations and orchards might promote natural regeneration of these sites.

ACKNOWLEDGEMENTS

This study was supported by the National Science Foundation of China (Grant No. 31040015), the Knowledge Innovation Project of Chinese Academy of Sciences (Grant No. KSCX2-YW-N-066) and West Light Foundation of Chinese Academy of Sciences. We are grateful to the staff of Xishuangbanna Station of Tropical Rainforest Ecosystem Studies for assistance in the field. We thank JY Cui and GD Tao for seedling identification, ZG Chen for assistance in the field and maintenance of seed germination experiments and LX Lin for data analysis.

REFERENCES

- AIDE TM & CAVELIER J. 1994. Barriers to lowland tropical forest restoration in the Sierra Nevada de Santa Marta, Colombia. *Restoration Ecology* 2: 219–229.
- ALVAREZ-AQUINO C, WILLIAMS-LINERA G & NEWTON AC. 2005. Disturbance effects on the seed bank of Mexican cloud forest fragments. *Biotropica* 37: 337–342.
- BLISS D & SMITH H. 1985. Penetration of light into soil and its role in the control of seed germination. *Plant Cell and Environment* 8: 475–483.
- CALLAWAY RM & MARON JL. 2006. What have exotic plant invasions taught us over the past 20 years? *Trends in Ecology and Evolution* 21: 369–374.
- CAO M, TANG Y, SHENG CY & ZHANG JH. 2000. Viable seeds buried in the tropical forest soils of Xishuangbanna, SW China. *Seed Science Research* 10: 255–264.
- CAO M & ZHANG JH. 1996. An ecological perspective on shifting cultivation in Xishuangbanna, SW China. *Wallaceana* 78: 21–27.
- CAO M, ZOU XM, WARREN M & ZHU H. 2006. Tropical forests of Xishuangbanna, China. *Biotropica* 38: 306–309.
- DALLING JW & BROWN TA. 2009. Long-term persistence of pioneer species in tropical rain forest soil seed banks. *American Naturalist* 173: 531–535.
- DALLING JW, SWAINE MD & GARWOOD NC. 1998. Dispersal patterns and seed bank dynamics of pioneer trees in moist tropical forest. *Ecology* 79: 564–578.
- DEROUW A. 1991. The invasion of *Chromolaena odorata* (L.) King and Robinson (Ex *Eupatorium odoratum*), and competition with the native flora, in a rainforest

zone, south-west Ivory Coast. *Journal of Biogeography* 18: 13–23.

- FLINT EP. 1994. Changes in land use in South and Southeast Asia from 1880 to 1980: a data base prepared as part of a coordinated research program on carbon fluxes in the tropics. *Chemosphere* 29: 1015–1062.
- FUNK JL & VITOUSEK PM. 2007. Resource-use efficiency and plant invasion in low-resource systems. *Nature* 446: 1079–1081.
- GUEVARA S, MORENO-CASASOLA P & SANCHEZ-RIOS G. 2005. Soil seed banks in the tropical agricultural fields of Los Tuxtlas, Mexico. *Tropical Ecology* 46: 219–227.
- HOLL KD, LOIK ME, LIN EHV & SAMUELS IA. 2000. Tropical montane forest restoration in Costa Rica: overcoming barriers to dispersal and establishment. *Restoration Ecology* 8: 339–349.
- HONU YAK & DANG QL. 2000. Responses of tree seedlings to the removal of *Chromolaena odorata* Linn. in a degraded forest in Ghana. *Forest Ecology and Management* 137: 75–82.
- HOOPER E, LEGENDRE P & CONDIT R. 2005. Barriers to forest regeneration of deforested and abandoned land in Panama. *Journal of Applied Ecology* 42: 1165–1174.
- LAURANCE WF. 2007. Forest destruction in tropical Asia. Current Science 93: 1544–1550.
- LEMENIH M & TEKETAY D. 2006. Changes in soil seed bank composition and density following deforestation and subsequent cultivation of a tropical dry Afromontane forest in Ethiopia. *Tropical Ecology* 47: 1–12.
- LI HM, AIDE TM, MA YX, LIU WJ & CAO M. 2007. Demand for rubber is causing the loss of high diversity rain forest in SW China. *Biodiversity and Conservation* 16: 1731–1745.
- LI HM, MA YX, LIU WJ & LIU WJ. 2009. Clearance and fragmentation of tropical rain forest in Xishuangbanna, SW China. *Biodiversity and Conservation* 18: 3421–3440.
- LIN LX, CAO M, HE YT, BASKIN JM & BASKIN CC. 2006. Nonconstituent species in soil seed banks as indicators of anthropogenic disturbance in forest fragments. *Canadian Journal of Forest Research* 36: 2300–2316.
- LIU WJ, MENG FR, ZHANG YP, LIU YH & LI HM. 2004. Water input from fog drip in the tropical seasonal rain forest of Xishuangbanna, South-West China. *Journal of Tropical Ecology* 20: 517–524.
- LIU WJ, ZHANG YP, LI HM & LIU YH. 2005. Fog drip and its relation to groundwater in the tropical seasonal rain forest of Xishuangbanna, Southwest China: a preliminary study. *Water Research* 39: 787–794.
- MACKENZIE DD & NAETH MA. 2010. The role of the forest soil propagule bank in assisted natural recovery after oil sands mining. *Restoration Ecology* 18: 418–427.
- MAMEDE MDA & DE ARAÚJO FS. 2008. Effects of slash and burn practices on a soil seed bank of Caatinga vegetation in northeastern Brazil. *Journal of Arid Environments* 72: 458–470.
- METCALFE DJ & TURNER IM. 1998. Soil seed bank from lowland rain forest in Singapore: canopy-gap and litter-gap demanders. *Journal of Tropical Ecology* 14: 103–108.
- OKE SO, AYANWALE TO & ISOLA OA. 2007. Soil seedbank in four contrasting plantations in Ile-Ife area of southwestern Nigeria. *Research Journal of Botany* 2: 13–22.

- PARROTTA JA & KNOWLES OH. 2001. Restoring tropical forests on lands mined for bauxite: examples from the Brazilian Amazon. *Ecological Engineering* 17: 219–239.
- PEARSON TRH, BURSLEM DFRP, MULLINS CE & DALLING JW. 2002. Germination ecology of neotropical pioneers: interacting effects of environmental conditions and seed size. *Ecology* 83: 2798–2807.
- PLUE J, GILS VAN B, PEPPLER-LISBACH C, SCHRIJVER DE A, VERHEYEN K & HERMY M. 2010. Seed-bank convergence under different tree species during forest development. *Perspectives in Plant Ecology Evolution and Systematics* 12: 211–218.
- QUINTANA-ASCENCIO PF, GONZALEZ-ESPINOSA M, RAMIREZ-MARCIAL N, DOMINGUEZ-VÁZQUEZ G & MARTINEZ-ICO M. 1996. Soil seed banks and regeneration of tropical rain forest from milpa fields at the Silva Lacandona, Chiapas, Mexico. *Biotropica* 28: 192–209.
- SIMPSON RL, LECK MA & PARKER VT. 1989. Seed banks: general concepts and methodological issues. Pp 3–8 in Leck MA, Parker VT & Simpson RL (eds) *Ecology of Soil Seed Banks*. Academic Press, San Diego.
- SINGHAKUMARA BMP, UDUPORUWA RSJP & ASHTON PMS. 2000. Soil seed banks in relation to light and topographic position of a hill dipterocarp forest in Sri Lanka. *Biotropica* 32: 190–196.
- SODHI NS, KOH LP, BROOK BW & NG PKL. 2004. Southeast Asian biodiversity: an impending disaster. *Trends in Ecology and Evolution* 19: 654–660.
- STANDISH RJ, CRAMER VA, WILD SL & HOBBS RJ. 2007. Seed dispersal and recruitment limitation are barriers to native recolonization of old fields in Western Australia. *Journal of Applied Ecology* 44: 435–445.
- SWAINE MD & WHITMORE TC. 1988. On the definition of ecological species groups in tropical rain forests. *Vegetatio* 75: 81–86.

- TANG Y, CAO M & FU XF. 2006. Soil seed bank in a dipterocarp rain forest in Xishuangbanna, Southwest China. *Biotropica* 38: 328–333.
- TILMAN D, FARGIONE J, WOLFF B, D'ANTONIO C, DOBSON A, HOWARTH R, SCHINDLER D, SCHLESINGER WH, SIMBERLOFF D & SWACKHAMER D. 2001. Forecasting agriculturally driven global environmental change. *Science* 292: 281–284.
- VÁZQUEZ-YANES C, ROJAS-ARÉCHIGA M, SÁNCHEZ-CORONADO ME & OROZCO-SEGOVIA A. 1996. Comparison of lightregulated seed germination in *Ficus* spp. and *Cecropia obtusifolia*: ecological implications. *Tree Physiology* 16: 871–875.
- WHITMORE TC. 1990. An Introduction to Tropical Rain Forests. Clarendon Press, Oxford.
- WIJDEVEN SMJ & KUZEE ME. 2000. Seed availability as a limiting factor in forest recovery processes in Costa Rica. *Restoration Ecology* 8: 414–424.
- YOUNG KR, EWEL JJ & BROWN BJ. 1987. Seed dynamics during forest succession in Costa Rica. *Vegetatio* 71: 157–173.
- ZHANG JH & CAO M. 1995. Tropical forest vegetation of Xishuangbanna, SW China and its secondary changes, with special reference to some problems in local nature conservation. *Biological Conservation* 73: 229–238.
- ZHANG ZQ, SHU WS, LAN CY & WONG MH. 2001. Soil seed bank as an input of seed source in revegetation of lead/zinc mine tailings. *Restoration Ecology* 9: 378–385.
- ZHU H, CAO M & HU HB. 2006. Geological history, flora, and vegetation of Xishuangbanna, southern Yunnan, China. *Biotropica* 38: 310–317.
- ZIMMERMAN JK, PASCARELLA JB & AIDE TM. 2000. Barriers to forest regeneration in an abandoned pasture in Puerto Rico. *Restoration Ecology* 8: 350–360.